



# Abschlussbericht

## Anschubprojekt: "Phosphorus as a cue regulating microbial N<sub>2</sub>O production"

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Vorhabensbezeichnung:	PQ4N
Laufzeit des Vorhabens:	15.09.2019-31.12.2019
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#### 1. Zusammenfassung und Schlussfolgerung

Phosphorus (P) has been shown to interact with nitrogen (N) transformations in soils, altering microbial sources of nitrous oxide ( $N_2O$ ) emissions. However, this P-regulated N response remains largely unclear. Interactions with water content as well as P-fertilisation history have rarely been investigated. Within this project, we carried out an incubation and a mesocosm experiment to increase understanding of the interactions.

Using a <sup>15</sup>N-labelled NO<sub>3</sub><sup>-</sup> tracer in a soil incubation experiment, we studied the influence of P addition on N conversions and N<sub>2</sub>O emission under soil water-holding capacities (WHC) of 45 and 60 %. We conclude from the results that P availability could increase N conversion via mineralization, decrease nitrification and increase denitrification, coupled with an increase in N<sub>2</sub>O production from denitrification, which was the main N<sub>2</sub>O source here despite moderate water contents.

In the mesocosm experiment, we studied effects of P-fertilisation history on reactions of N<sub>2</sub>O production processes to P addition. Therefore, soils from two treatments of a long-term P fertilisation experiment were incubated with or without P addition. Fluxes of NO, N<sub>2</sub>O and CO<sub>2</sub> were measured continuously and event-based samples taken for isotopomeric measurement of N<sub>2</sub>O and molecular analysis of the microbial community. While long-term P fertilisation decreased N<sub>2</sub>O production, short-term P addition increased it, leading to largest cumulative N<sub>2</sub>O production from the low-P soil with P fertilisation. Preliminary isotopic signatures did not suggest differences in N<sub>2</sub>O sources among treatments, but over time of incubation. Molecular results are pending due to constraints caused by pandemic measures. Overall, the seed project has shown that P has an effect on N conversions and N<sub>2</sub>O production. This effect depends on P fertilisation history and probably on the microbial community. These findings will be published in two papers. A grant proposal on this topic seems promising after submitting the manuscripts.

#### 2. Einleitung und Ziele des Projektes

Agricultural soils are an important source of the greenhouse gas nitrous oxide (N<sub>2</sub>O). N<sub>2</sub>O is mainly produced by microbial transformations of N in soils and is often enhanced where available N exceeds crop demand (Oenema et al., 2005). Different pathways, e.g. denitrification, nitrification, or nitrifier denitrification, are involved in the production of N<sub>2</sub>O under a range of soil conditions (Wrage-Mönnig et al., 2018). Among environmental factors, N availability and soil water content are major drivers of N<sub>2</sub>O emissions (Butterbach-Bahl et al. 2013; Chen et al., 2014). However, there are also interactions with other elements, e.g. carbon (C) and phosphorus (P) (He and Djikstra, 2015; O'Neill et al., 2020).

Like N, phosphorus (P) is a critical nutrient for plant and microbial growth. P is a primary constituent of biomolecules such as nucleic acids (DNA and RNA), phospholipids and





adenosine triphosphate (ATP), which regulate many biological processes including energy transfer reactions and activation of enzymes (Nannipieri and Paul, 2009; Rouached et al., 2010). Availability of P is essential for the metabolisms of carbon and amino acids, e.g. during photosynthesis and respiration in plants and microbes (Elanchezhian et al., 2015), which play an important role in regulating N cycling and availability. Besides, N transformations in the soil are multi-enzyme processes carried out by diverse microbial species. The effects of P on various enzyme activities and composition of the microbial community could differ, leading to different P sensitivity of e.g. N mineralization, nitrification and denitrification (Olander and Vitousek, 2010).

The N-regulated P response in plant and microbial activity has been well described by recent studies (Medici et al., 2019; Hu et al., 2019), whereas the P-regulated N response remains largely unclear (Hu and Chu, 2019). So far, no robust information is available for P effects and P fertilisation history on sources of N<sub>2</sub>O emission. Therefore, in this seed project, we used isotopic approaches to improve knowledge on N<sub>2</sub>O sources and N transformations after addition of P fertiliser. In an incubation experiment, interactions between water-holding capacity and P addition were studied. In a mesocosm experiment, we used soils with different histories of P-fertilisation. We hypothesized that 1) P addition would decrease inorganic N in soil and N<sub>2</sub>O emissions due to stimulated biological uptake, 2) there would be an interaction between water content and P addition on N<sub>2</sub>O emissions, 3) P addition would have no effect on the relative importance of nitrification or denitrification for N conversion or as sources of N<sub>2</sub>O and 4) P addition effects would depend on the history of P fertilisation.

#### 3. Material und Methoden

#### Set-up of the incubation experiment

The soil was obtained from a low-input agricultural system from the experimental station of the University of Rostock, managed in order to decrease soil nutrient levels for experimental purposes. For P, the soil had been classified in the German fertilizer content class A (very low) for the last years. 200 g air-dried soil sieved over 2 mm was placed into 750 mL glass jars with the following four treatment combinations (n = 3): with or without phosphorus (P) at 45 % or 60 % WHC. Jars were divided into two groups, one for gas flux and isotopic gas measurements, and the other for measuring mineral N concentrations and <sup>15</sup>N signatures of NO<sub>3</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup>. Soils were pre-incubated for two days at 30 % WHC, after which <sup>15</sup>N-NO<sub>3</sub><sup>-</sup>-labelled NH<sub>4</sub>NO<sub>3</sub> tracer (10 atom%) at a rate of 50 mg N kg<sup>-1</sup> were stirred into the soil in all the glass jars. Half of the jars received P (triple super phosphate; 225 mg P kg<sup>-1</sup> soil) after adjusting to the target soil water content, while the other half received no P addition. The targeted soil water content was maintained daily on a weight basis. The jars were closed for 1 h with air-tight rubber lids with septa before taking the gas samples.





Gas samples were taken at 24 h intervals from the headspace, transferred into heliumflushed and evacuated 12 mL exetainer vials and analyzed for N<sub>2</sub>O and its isotopic signatures using a trace gas preparation unit (Trace Gas, Elementar, UK) coupled to an isotope ratio mass spectrometer (IRMS, Isoprime 100, Elementar, GermanyUK). Fluxes were calculated by assuming a linear relationship between concentration and incubation time. For calibration, two working standards were used that had been calibrated against the standards of the laboratory of the Department of Environmental System Science, ETH Zürich (Verhoeven et al. 2019). Sample peak ratios are initially calibrated against an N<sub>2</sub>O reference gas peak (100 % N<sub>2</sub>O, Air Liquide, Germany) and then corrected for drift and span using the working standards.

The second group of jars was used to extract NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> on days 1, 3 and 5 of the incubation by adding 150 mL of 1 M KCl solution to 40 g subsamples of the treated soils, shaking for 1 hour and filtering. The concentrations and <sup>15</sup>N enrichment of NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> were determined by micro-diffusion (Brooks et al., 1989) and analysis on an elemental analyzer (vario PYRO cube, Elementar, Germany) interfaced to the above IRMS. Our internal standards (wheat flour and sulfanilamide) were calibrated against IAEA-600 and IAEA-NO-3 for <sup>15</sup>N. All isotopic values for this experiment are given in at% excess.

#### Set-up of the mesocosm experiment

The mesocosm experiment was carried out at the facilities of the University of Goettingen, Plant Nutrition and Yield Physiology. Soil was taken from two treatments of the Rostock long-term field experiment (Zicker et al., 2018): a control treatment without added P since 1998 (low P soil) and a treatment with cattle manure + TSP (high P soil; manure applied every three years at about 30 t ha<sup>-1</sup> and TSP annually at 21.8 kg P ha<sup>-1</sup> until 2013 and 30 kg P ha<sup>-1</sup> thereafter). The soils were packed into the mesocosms at a bulk density of 1.46 g cm<sup>-3</sup>. There were four treatments (five replicates each): low P soil without added P (LP-), low P soil with added P (LP+), high P soil without added P (HP-) and high P soil with added P (HP+). P was added as TSP at 34.4 mg P kg<sup>-1</sup>, which was supposed to bring the low P soil to the P content of the high P soil. All treatments received 50 mg N kg<sup>-1</sup> as NH<sub>4</sub>NO<sub>3</sub> at the start of incubation and moisture was kept at 75% WHC. To check for potential carbon (C) limitation, on the 6<sup>th</sup> day of incubation, both glucose (300 mg C kg<sup>-1</sup>) and NH<sub>4</sub>NO<sub>3</sub> (50 mg N kg<sup>-1</sup>) were added to all treatments.

Fluxes of N<sub>2</sub>O, nitric oxide (NO) and carbon dioxide (CO<sub>2</sub>) were measured continuously over 13 days. Furthermore, air samples were taken from the pots event-based on days 1, 2, 3, 7, 10 and 13 and measured for stable isotope composition and site preference on the IRMS as above. Isotopic signatures are here reported as  $\delta$ -values. Samples for soil pH, NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> (KCl extraction for concentrations and isotopic composition) and inorganic P content (CAL





extraction) were taken on days 0, 3, 6, 9 and 13. Part of these samples was frozen immediately for later molecular analyses. Due to the outbreak of Covid-19 and following lock-down, sample transfer had to be delayed and molecular analyses will be conducted in the coming weeks.

#### Statistical analyses

Normality and homogeneity of variances for all variables was verified using the Kolmogorov-Smirnov test before further statistical analysis. In the incubation experiment, a repeated measures analysis of variance (ANOVA) was used to test for main effects of P, WHC, day of measurement, and their interactions. Two-way ANOVA was used to test for P, WHC, and their interaction on cumulative N<sub>2</sub>O in soil. Where treatment effects were significant at P < 0.05, least significant difference (LSD) tests were used to compare the means of each treatment combination. P values between 0.05 and 0.10 were considered as marginally significant. All analyses were performed with the R software (version 3.6.1) for Windows.

#### 4. Ergebnisse

#### Incubation experiment

Analysis of the experimental soil before and after incubation showed that the concentrations of  $NO_3^-$  were larger than those of  $NH_4^+$ . Overall, the  $NH_4^+$  concentrations increased significantly over time (P = 0.030), whereas  $NO_3^-$  decreased (significant day effect, P < 0.001) (Fig. 1a). Both the NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> pools were not significantly affected by P treatment. Soil water content significantly (P < 0.001) affected the NO<sub>3</sub><sup>-</sup> concentration, which was higher at 45 % WHC than at 60%. The  $NH_4^+$  concentration did not change with WHC. There was a significant soil water content × day interactive effect (P = 0.020) in NO<sub>3</sub><sup>-</sup>, which was larger on day 3 compared to day 5 (Fig. 1a). Overall, the NO<sub>3</sub><sup>-</sup> concentration was marginally significantly (P = 0.070) increased without  $P \times 45$  % WHC compared to the other treatments. The <sup>15</sup>N-enrichment of the  $NH_4^+$  pool slightly increased over time (significant day effect, P = 0.001), while that of the NO<sub>3</sub><sup>-</sup> pool significantly decreased (P < 0.001) (Fig. 1b). P addition did not affect the <sup>15</sup>N signature of the NH<sub>4</sub><sup>+</sup> pool, which remained at background levels (0.011 – 0.013 at% excess; P = 0.003). Overall, the NO<sub>3</sub><sup>-</sup> pool was significantly (P = 0.001) more enriched with P addition (2.56 at% excess) than without (2.44 at% excess). With a soil water content of 60 % WHC, the <sup>15</sup>N-enrichment of the NH<sub>4</sub><sup>+</sup> pool was slightly increased compared to 45 % (significant WHC effect, P = 0.003), while that of  $NO_3^-$  remained the same. There was, however, a trend to a significant P  $\times$  soil water content interactive effect for the <sup>15</sup>N- $NO_3^-$  enrichment (P = 0.050), which was larger with P × 45 % WHC compared to the other treatments.

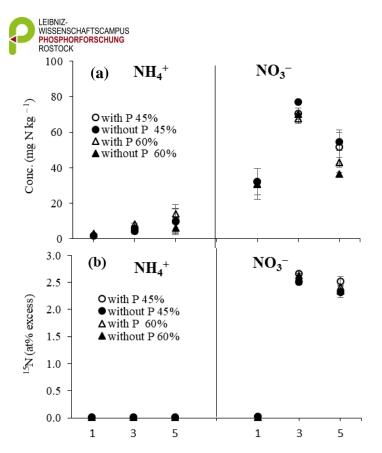




Fig. 1. Mean  $NH_4^+$  and  $NO_3^-$  concentrations (a) and the <sup>15</sup>N-enrichments (b) as affected by P and soil water content during the 5-day incubation. Error bars represent standard deviation. Data for day 1 were background values taken before the addition of P or <sup>15</sup>N-NO<sub>3</sub><sup>-</sup>.

 $N_2O$  emission and its <sup>15</sup>N-enrichment were largest on day 2 (Fig. 2a, b) (significant day effect, P < 0.0001, for both N<sub>2</sub>O and <sup>15</sup>N-N<sub>2</sub>O). The addition of P increased the emission of N<sub>2</sub>O, which was marginally significant (P = 0.080). WHC did not show any significant effect on N<sub>2</sub>O emission. The <sup>15</sup>N-enrichment of N<sub>2</sub>O (Fig. 2b) and cumulative N<sub>2</sub>O emission were not significantly different among the treatments. Also, there were no significant interactions between P and WHC for N<sub>2</sub>O emission or its enrichment.

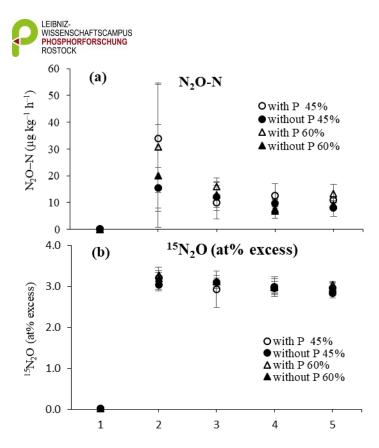




Fig. 2. Mean N<sub>2</sub>O–N fluxes (a) and the <sup>15</sup>N-enrichment (b) as affected by P and soil water content during the 5-day incubation. Error bars represent standard deviation. Data for day 1 were background values taken before the addition of P or <sup>15</sup>N-NO<sub>3</sub><sup>-</sup>

Mesocosm experiment (preliminary results)

The soil P content was clearly different in the different treatments: P fertilisation had an influence on soil P, but did not raise the inorganic P content in the LP+ soil to that of the unfertilised HP- soil (data not shown).

Fluxes of N<sub>2</sub>O, NO and CO<sub>2</sub> were dynamic over time (Fig. 3). For N<sub>2</sub>O, there was a peak emission event at the beginning of the incubation. This was larger and the maximum a little later in the LP than in the HP soils and was slightly increased by P fertilisation in both soils. The addition of glucose and N on the sixth day of incubation led to small peaks of N<sub>2</sub>O, especially in the LP+ soil. Furthermore, large NO fluxes were measured afterwards. A first NO peak was similar for all treatments, but it was followed by a second peak in N<sub>2</sub>O and NO that was much smaller and a little quicker in both HP incubations. There was also an increased CO<sub>2</sub> flux at this time, which had a similar magnitude for all treatments. Accumulated N<sub>2</sub>O emissions (Fig. 4) were slightly increased with P fertilisation, but were larger in LP than HP soils.





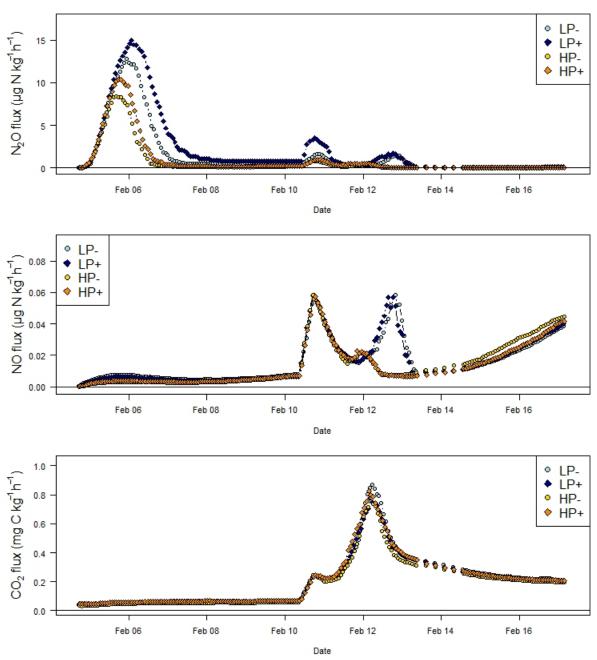
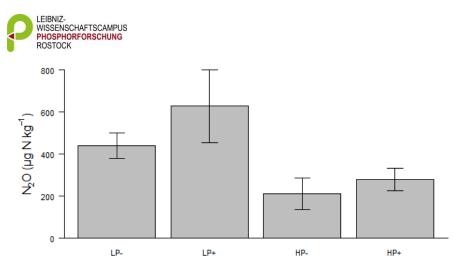
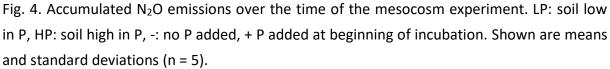


Fig. 3. Fluxes of  $N_2O$ , NO and  $CO_2$  over time in treatments differing in P fertilisation. LP: soil low in P, HP: soil high in P, -: no P added, + P added at beginning of incubation. Shown are means (n = 5).





The isotopic signatures of N<sub>2</sub>O (Fig. 5) showed large variability over time. The <sup>15</sup>N signatures were depleted at the beginning of the incubation and became gradually more enriched towards the end. In contrast,  $\delta^{18}$ O signatures were rather depleted on the first day, at around 40 ‰ on the next three sampling days and at about 25 ‰ towards the end of the incubation. The site preference was negative throughout, but especially on days 3, 7 and 10, indicating that the N at the  $\alpha$ -position was more depleted than that at the  $\beta$ -position. There were no consistent differences in isotopic signatures among treatments.

#### 5. Diskussion

#### Incubation experiment

In the following, we will discuss the results according to our hypotheses. We will start with effects of P addition on inorganic N in soil and N<sub>2</sub>O emissions, go on with the influence of the interaction between water content and P addition on N<sub>2</sub>O emissions, and lastly discuss effects on the relative importance of nitrification or denitrification for N conversion or as sources of N<sub>2</sub>O.

Our results did not support the hypothesis that P addition would decrease inorganic N in soil and N<sub>2</sub>O emissions by stimulated biological uptake. The addition of P did not affect  $NO_3^$ availability, though P addition appeared to increase both the NH<sub>4</sub><sup>+</sup> concentration and N<sub>2</sub>O emission. A similar P addition effect increasing N<sub>2</sub>O production has been found in a P-limited grassland soil; however, extractable NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup> were not significantly affected by the P addition (He and Dijkstra, 2015; Mehnaz and Dijkstra, 2016). Previous studies have demonstrated that alleviation of P limitation may increase or decrease N<sub>2</sub>O emission by stimulating biological N uptake (Baral et al., 2014; He and Dijkstra, 2015; Mehnaz and Dijkstra, 2016). Our study did not include plant activity and responses, which might explain a





missing effect of P on inorganic N and N<sub>2</sub>O emissions by biological N uptake. Baral et al. (2014) demonstrated that P addition decreased N<sub>2</sub>O emissions by increasing plant N uptake in a P-limited arable soil with maize under greenhouse conditions. Conversely, He and Dijkstra (2015) found that P addition did not increase plant N uptake but instead increased N loss in P-limited soil in a mesocosm experiment with grass species. Thus, it seems that if P addition leads to increased plant N uptake, N<sub>2</sub>O emissions are decreased, and vice versa.

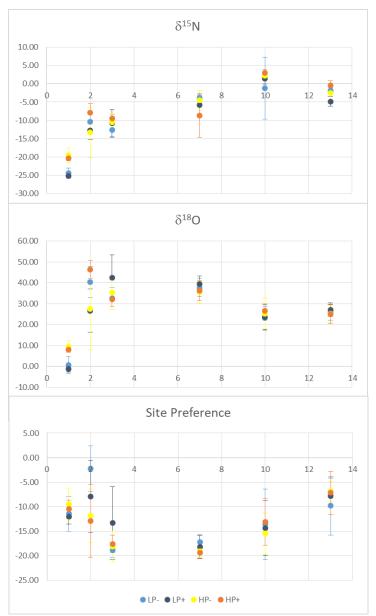


Fig. 5. Isotopic signatures of N<sub>2</sub>O over the course of the mesocosm experiment. Shown are means and standard deviations (n = 5). All values are  $\delta$ -values in ‰.

Our results did not support the second hypothesis of an interaction between soil water content and P addition on N<sub>2</sub>O emissions. However, the soil water content influenced  $NO_3^-$  availability. The slight increase in  $NO_3^-$  concentrations without P at 45 % WHC compared to





the other treatments suggests that  $NO_3^-$  reduction to  $N_2O$  was affected by the interaction.  $N_2O$  emissions were smaller at peak emission and decreased more until day 5 without P under 45 % WHC than 60 %, indicating increased reduction of  $NO_3^-$  to  $N_2O$  under the higher soil water content. Decreased  $NO_3^-$  concentrations under the higher soil water content are also indicative of increased  $NO_3^-$  use relative to production and thus in our plantless system potentially larger denitrification relative to nitrification. An increasing water content is known to increase denitrification relative to nitrification and at very large water contents increase  $N_2O$  reduction (Butterbach-Bahl et al. 2013).

Our hypothesis that P addition would have no effect on the relative importance of nitrification or denitrification for N conversion or as sources of N<sub>2</sub>O was not supported. N transformations and the associated N<sub>2</sub>O emissions were stimulated after P addition. Our result is more consistent with a larger increase of mineralization and denitrification compared to nitrification with P addition, which suggests that P availability could regulate inorganic N conversion and associated N<sub>2</sub>O production. An increased NH<sub>4</sub><sup>+</sup> concentration with P addition indicated increasing mineralization relative to nitrification. Moreover, the isotopic signatures of the NH<sub>4</sub><sup>+</sup> pool after addition of the <sup>15</sup>NO<sub>3</sub><sup>-</sup> label indicated no occurrence of dissimilatory nitrate reduction to ammonia (DNRA) in this soil. A tendency to more depleted <sup>15</sup>N-NH<sub>4</sub><sup>+</sup> signatures with added P hints towards a decrease of nitrification relative to mineralization with P.

Although the overall NO<sub>3</sub><sup>-</sup> concentrations were not affected by P treatment alone in our soil incubation experiment, decreased NO<sub>3</sub><sup>-</sup> concentrations over time and significantly increased <sup>15</sup>N-NO<sub>3</sub><sup>-</sup> with added P showed increasing denitrification relative to nitrification. The decline of the <sup>15</sup>N-NO<sub>3</sub><sup>-</sup> enrichment over time indicates dilution by unlabelled NO<sub>3</sub><sup>-</sup> produced from nitrification. As the enrichment remains larger with P addition than without, the influence of nitrification seems decreased with P. Interestingly, the <sup>15</sup>N-enrichment of N<sub>2</sub>O almost matched that of NO<sub>3</sub><sup>-</sup>, indicating that mostly denitrification was responsible for N<sub>2</sub>O production, despite the rather dry soil conditions. Similar to our findings, P addition also increased N<sub>2</sub>O production by denitrification compared to nitrification where NO<sub>3</sub><sup>-</sup> was abundant in a highly P-limited grassland soil (Mehnaz and Dijkstra, 2016). To summarise, our soil incubation experiment indicates that P addition enhanced N conversion via mineralization and N<sub>2</sub>O emissions via denitrification, whereas it decreased nitrification relative to the other processes.

#### Mesocosm experiment

The preliminary results of the mesocosm experiment showed that the effect of P addition on  $N_2O$  production indeed depended on the soil's history of P fertilisation. Interestingly, short-term P addition had a stimulating effect on  $N_2O$  emission, while long-term P-fertilisation had





a reducing effect. Isotopic results did not indicate different sources of  $N_2O$  among treatments. Here, signatures of  $^{15}N$  in soil  $NH_4^+$  and  $NO_3^-$  need to be considered for further evaluation. Besides, molecular analyses will indicate whether differences existed among the soils initially and/or after incubation.

#### 6. Weitere Leistungen und Ziele aus dem Projekt

At least two publications will be prepared from this project. One concerning the incubation experiment is almost ready to be submitted – this report was partly based on the text of the manuscript. The other manuscript will be written after submission of the first and analysis of the remaining data for <sup>15</sup>N (mineral N in soil) and samples for microbial composition. There are several ideas for research proposals.

#### 7. Literaturverzeichnis

Baral, B.R., Kuyper, T.W., Van Groenigen, J.W., 2014. Liebig's law of the minimum applied to a greenhouse gas: alleviation of P-limitation reduces soil N2O emission. Plant Soil. 374, 539–548. https://doi.org/10.1007/s11104-013-1913-8.

Brooks, P.D., Stark, J.M., McInteer, B.B., Preston, T. 1989. Diffusion method to prepare soil extracts for automated nitrogen-15 analysis. Soil Sci. Soc. Am. J. 53, 1707–1711. https://doi.org/10.2136/sssaj1989.03615995005300060016x.

Butterbach-Bahl, K., Baggs, E.M., Dannenmann, M., Kiese, R., Zechmeister-Boltenstern, S., 2013. Nitrous oxide emissions from soils: how well do we understand the processes and their controls? Phil. Trans. R. Soc. B.368 20130122. http://dx.doi.org/10.1098/rstb.2013.0122.

Chen, Z.M., Ding, W.X., Luo, Y.Q., Yu, H.H., Xu, Y.H., Müller, C., Xu, X., Zhu, T.B., 2014. Nitrous oxide emissions from cultivated black soil: a case study in Northeast China and global estimates using empirical model. Glob. Biogeochem. Cycles. 28, 1311–1326. https://doi.org/10.1002/2014GB004871.

Elanchezhian, R., Krishnapriya, V., Pandey, R., Rao, A.S., Abrol, Y.P., 2015. Physiological and molecular approaches for improving phosphorus uptake efficiency of crops. Curr. Sci. 108, 1271–1279.

He, M., Dijkstra, F.A., 2015. Phosphorus addition enhances loss of nitrogen in a phosphoruspoor soil. Soil Biol. Biochem. 82, 99–106. http://dx.doi.org/10.1016/j.soilbio.2014.12.015.

Hu B., Chu, C., 2019. Nitrogen-phosphorus interplay: old story with molecular tale. New Phytol. https://doi.org/10.1111/nph.16102.

Hu, B., Jiang, Z., Wang, W., Qiu, Y., Zhang, Z., Liu, Y., Li, A., Gao, X., Liu, L., Qian, Y., Huang, X., Yu, F., Kang, S., Wang, Y., Xie, J., Cao, S., Zhang, L., Wang, Y., Xie, Q., Kopriva, S., Chu, C., 2019.





Nitrate-NRT1.1B-SPX4 cascade integrates nitrogen and phosphorus signalling networks in plants. Nat. Plants. 5, 401–413. https://doi.org/10.1038/s41477-019-0384-1.

Medici, A., Szponarski, W., Dangeville, P., Safi, A., Dissanayake, I.M., Saenchai, C., Emanuel, A., Rubio, V., Lacombe, B., Ruffel, S., Tanurdzic, M., Rouached, H., Krouka, G., 2019. Identification of molecular integrators shows that nitrogen actively controls the phosphate starvation response in plants. The Plant Cell. 31, 1171–1184. https://doi.org/10.1105/tpc.18.00656.

Mehnaz, K.R., Dijkstra, F.A., 2016. Denitrification and associated  $N_2O$  emissions are limited by phosphorus availability in a grassland soil. Geoderma. 284:34–41. http://dx.doi.org/10.1016/j.geoderma.2016.08.011.

Nannipieri, P., Paul, E., 2009. The chemical and functional characterization of soil N and its biotic components. Soil Biol. Biochem. 41, 2357–2369. https://doi.org/10.1016/j.soilbio.2009.07.013.

O'Neill, R.M., Girkin, N.T., Krol, D., Wall, D.P., Brennan, F.P., Lanigan, G.L., Renou-Wilson, F., Müller, C., Richards, K.G., 2020. The effect of carbon availability on N<sub>2</sub>O emissions is moderated by soil phosphorus Soil Biol. Biochem. 142. 107726. https://doi.org/10.1016/j.soilbio.2020.107726.

Oenema, O., Wrage, N., Velthof, G.L., van Groenigen, J.W., Dolfing, J., Kuikman, P.J., 2005. Trends in global nitrous oxide emissions from animal production systems. Nutr. Cycl Agroecosys. 72, 51-65. https://doi.org/10.1007/s10705-004-7354-2.

Olander, L.P., Vitousek, P.M., 2000. Regulation of soil phosphatase and chitinase activity by N and P availability. Biogeochem. 49:175–190. https://doi.org/10.1023/A:1006316117817

Rouached, H., Arpat, A.B., Poirier, Y., 2010. Regulation of phosphate starvation responses in plants: signalling players and crosstalks. Mol. Plant. 3, 288–299. https://doi.org/10.1093/mp/ssp120.

Verhoeven, E., Barthel, M., Yu, L., Celi, L., Said-Pullicino, D., Sleutel, S., Lewicka-Szczebak, D., Six, J., Decock, C., 2019. Early season N<sub>2</sub>O emissions under variable water management in rice systems: source-partitioning emissions using isotope ratios along a depth profile. Biogeosciences 16:383-408. https://doi.org/10.5194/bg-16-383-2019.

Zicker, T., von Tucher, S., Kavka, M., Eichler-Löbermann, B., 2018. Soil tst phosphorus as affected by phosphorus budgets in two long-term field experiments in Germany. Field Crop Research 218:158-170. https://doi.org/10.1016/j.fcr.2018.01.008.

#### Danksagung

The authors wish to thank Pauline Rummel and Prof. Dr. Klaus Dittert, Goettingen University, for the possibility to use their mesocosm unit and for the superb cooperation. We thank





Diana Werner for brilliant technical assistance in the lab of Grassland and Fodder Sciences, University of Rostock.